



A simple guide to assessing maturity and biodiversity in Mediterranean forest stands



A SIMPLE GUIDE TO ASSESSING MATURITY AND BIODIVERSITY IN MEDITERRANEAN FOREST STANDS

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Cover photo: centenary oak with partially dead crown and woodpecker breeding cavities (Photo: Lluís Comas).

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1. INTRODUCTION

t is vital to assess forest habitats for biodiversity and maturity, not least because it is estimated that forests are home to more than two-thirds of the earth's biodiversity (WCFSD 1999). A significant part of this biodiversity is associated with the more mature stages of the forest (Wirth et al. 2009; Hilmers et al. 2018), highlighting the need for biodiversity assessment at each stage in the life cycle of a forest.

Biological diversity in forests results from evolutionary processes that are millions of years old and are driven by the network of interactions between species and disturbances. The conservation of biological diversity is fundamental for the maintenance of these ecological processes (FAO 2020). The increase in complexity of ecosystems throughout an ecological succession in forests is known as silvogenetic cycle and reaches its zenith in the later stages, with the greatest complexity and, consequently, the greatest biological diversity. (Kuusinen and Siitonen, 1998; Redecker et al. 2001; Jacobs et al. 2007; Avila-Cabadilla et al. 2009; de la Peña-Cuéllar et al. 2012; Hilmers et al. 2018). Mature and senescent stands contain a greater amount and diversity of resources, structures and micro-climates, facilitating the coexistence of multiple species, increasing the number of niches and reducing the risk of local extinction (Schowalter 1995; Ferris and Humphrey 1999; Stein and Kreft 2015). These stands, which are more stable ecosystems and resilient to disturbance, strengthen the resilience of adjacent forest areas with less biodiversity (Bauhus et al. 2017; Gustafsson et al. 2019). The presence in space and time of forests with all phases of the silvogenetic cycle creates great heterogeneity and high biodiversity in mature and senescent

stands, absent in the previous ones. These include species that have restricted distribution and which are highly vulnerable to man-made disturbances and most of them endangered (EUROPARC-Spain 2020a).

The main way to assess the biodiversity of a forest at stand scale is by exhaustively sampling taxonomic groups, known as bioindicators, that are sensitive to changes in the ecosystem and which, as a whole, are indicative of changes in the global biodiversity of the habitat. However, taxonomic inventories are time-consuming and costly, and require specialist experts. Another way to assess biodiversity is through the use of proxy indicators that are easier to monitor in the field, by identifying the plant and physical habitat structures on which taxonomic groups depend (e.g., Lindenmayer et al. 2000 and 2006). However, to date, it has not been possible to identify a complete list of key attributes or features that are fully valid for all taxonomic groups and all forest habitats. Gao et al. (2015) and Larrieu et al. (2019) found significant relationships between structural elements and certain taxonomic groups, particularly saproxylic beetles, followed by soil beetles, aphyllophorous fungi and mosses. Other studies point out the close correlation between a richness and abundance of forest birds (passerines and woodpeckers in particular) and structural elements at the stand scale (a dozen hectares upwards), reflecting the size of passerine nesting territories (Camprodon 2013). Bats respond better over a larger area, as their hunting ranges are much more extensive and they have a fairly close correlation to increasing forest maturity (Camprodon et al. 2010).



Figure 1. In Mediterranean forests woodpeckers excavate their nests in decaying and dead standing trees of diameter class 20 upwards. In the image, a finished cavity and a recently started cavity made by a great spotted woodpecker (photo: Jordi Bas).

In order to have an idea of the real biodiversity, it is essential, therefore, to periodically monitor certain taxonomic groups. But which should be measured? Ideally, we should track taxonomic or functional groups with the highest bioindicator value. Many taxonomic groups complement each other, so several must be monitored at the same time to gain a more complete picture of the real biodiversity. One of the best groups, due to their short life cycle and high taxonomic diversity (the greatest in the forest), is insects. They are essential to many ecosystem processes and functions, sensitive to changes in their environment. Another approach is the indirect assessment of certain key features, where there is clear evidence that said features are closely correlated with the presence of certain communities of organisms. For example, deadwood is closely correlated with saproxylic organisms. A stand can also have the capacity to host, for example, forest birds, because there are large trees and a heterogeneous structure. However, if it does not contain deadwood or specific tree microhabitats, then typical forest species such as saproxylic invertebrates and fungi will not be present.



Figure 2. Large lying deadwood in different stages of decomposition (photo: Lluís Comas).

Finally, since there is a close link between high biological diversity and forest maturity, it is possible to define a series of attributes associated with these processes. These attributes are the result of forest dynamics that act over hundreds of years, allowing natural processes to take place over time, provided no major natural or man-made disturbances occur. The key factors at the stand scale are:

- The spatial heterogeneity of the forest: small open areas exposed to the sun alongside shady, damper areas, a mix of soil types, lithology, topography and altitudinal gradients, providing a greater number of ecological niches.
- The diversity of plant species. The presence of different species of trees, shrubs and

herbaceous plants provides a wide range of trophic resources.

- The structural complexity of the forest. The presence of trees of different sizes and ages, as well as fallen trunks and other structural elements, creates microhabitats and refuges.
- *The abundance of deadwood*, of any size, lying or standing, and different stages of decay. Thousands of saproxylic species depend on this resource, in many cases exclusively.
- Species interaction. Interactions between organisms over time and in space, and the functions they perform are essential to maintain the diversity, health and productivity of the forest ecosystem.



Figure 3. The spatial heterogeneity of the forest may also reflect the diversity of soil characteristics, lithology and/or topography (photo: Lluís Comas).

2. PURPOSE OF THE GUIDE

he main purpose of this guide is to present a methodology for diagnosing **maturity** and biodiversity hosting potential at stand scale using direct and indirect indicators

for Mediterranean forest habitats. This guide defines and explains the indicators used, the thresholds for assessing them and the common field methodology used to carry out said diagnosis.

3. BIODIVERSITY AND MATURITY ASSESSMENT SYSTEM

his common guide details two well-established methodologies with a high degree of consensus that are used to simultaneously assess forest biodiversity and maturity: the **Index of Biodiversity Potential (IBP)** and the **Redbosques Maturity Index**. The IBP was designed and tested to assess a stand's capacity to host forest taxa (animals, plants and fungi), based on structural and context indicators (Gonin et al. 2012). The RedBosques methodology was developed as part of the Life-RedBosques project

(EUROPARC-Spain, 2020b) based on work carried out by Rossi and Vallauri (2013). The index evaluates the forest's naturalness, i.e. its maturity, human footprint and spatial integrity. Although both indices use similar indicators for structure and composition, there are some differences in how they are defined and in the sampling methodology. This guide describes the indicators used in both methodologies, their differences, and proposes a common field sampling methodology for performing a combined assessment.



Figure 4. Girdling a stone pine as part of a LIFE BIORGEST natural dynamics measure (photo: Jordi Camprodon).

4. STAND INDICATORS

he two indices have different aims: while the RedBosques Maturity Index (RB) measures a stand's maturity, the IBP indirectly estimates the potential taxonomic diversity that a stand can host. As the sampling methods and how the variables are measured in the field vary

slightly, the similarities and differences for each indicator are detailed in the section below. The reason for choosing each indicator is also explained. Table 1 summarises all the indicators and Table 2 sets out the thresholds for classifying a stand in terms of potential biodiversity or maturity.

TABLE 1.

Comparison of the main methodological differences and sampling constraints for the two assessment protocols: RB (Redbosques Maturity Index), IBP (Biodiversity Potential Index). The sampling details and constraints for each protocol are contained in the guide.

Indicator	Protocol	Scale	Description	Differences / Constraints
Native tree species	RB	Stand	Number of different native tree species at any stage of develop- ment present in the stand	Live h≥50 cm
	BP	Stand	Factor A. Number of genera other than native tree species at any stage of development, dead or live, present in an area of 1 hectare.	Dead or live h≥50 cm
Basal area	sal area RB Plot Average basal area (m²/ trees of DBH > 17.5 cm) o plots		Average basal area (m²/ha) (live trees of DBH > 17.5 cm) of all the plots	DBH≥17.5 cm
	IBP Not used in assessment		Not used in assessment	
Vertical strata	RB	Plot	Number of strata. There are four strata of equal height (tree species only, at any stage of development) + 1 emergent stratum	CC ≥ 20%
	BP	Stand	Factor B. Number of strata - 1 herbaceous and semi-woody stratum - 4 woody strata: very low (< 1.5 m); low (1.5-5 m); intermediate (5-15 m) and tall (≥ 15 m)	CC ≥ 20%

Indicator	Protocol	Scale	Description	Differences / Constraints
Diametric classes	RB	Stand	Number of DCs other than native tree species present in all the plots sampled	DBH≥17.5 cm
	IBP		Not used in assessment	
Large and very large trees	RB	Plot	Number of exceptional live trees per hectare. A tree is considered exceptional if its DBH in cm is at least three times the dominant height in m (H_0) of the species in the stand.	DBH≥3 x Ho
	IBP	Plot	Factor E. Number of live trees per hectare of: - Large trees (LT) - Very large trees (VLT)	- LT (37.5 <dbh<57.5 cm)<br="">- VLT (DBH≥57.5 cm) or (DBH≥37.5 cm)*</dbh<57.5>
Medium and large deadwood	RB	Stand	Volume of standing or lying dead- wood of any tree species Percentage (%) of total deadwood volume (standing and lying) in re- lation to the volume of live trees	DBH≥17.5 cm
	(BP)	Plot	Factor C. Standing dead trees or snags of Medium Deadwood (MDW) and/or Large Deadwood (LDW) at least 1 metre high (H) Factor D. Lying medium dead- wood (MDW) and/or large dead- wood (LDW) of at least 1 meter length (L)	- H o L≥1 m - MDW (17.5 <dbh<27.5 cm)<br="">- LDW (DBH>27.5 cm) or (DBH≥17.5 cm)*</dbh<27.5>
Tree microhabitats (TreM)	RB	Stand	Number of different types of TreM detected in all the plots (based on the 10 proposed types). A TreM type counts if there are at least two per hectare.	
	IBP	Stand	Factor F. Number <u>of live trees with</u> <u>TreMs</u> per hectare (record and classify, based on the 15 types, all trees with TreMs observed up to a maximum of two trees/ha × TreM group).	

* For site quality type C (poor) or for species in slow-growing genera (Arbutus, Acer, Pyrus, Sorbus, etc.)

Indicator	Protocol Scale		Description	Differences / Constraints		
Flower-rich open areas	RB		Not used in assessment			
	(BP)	Plot	Factor G. Percentage (%) of surface area containing open spaces with flowering vegetation			
Dynamic	RB Stand		Each phase of the forest dynamics cycle is represented in the stand (1. Gap, 2. Regeneration, 3. Oc- cupation, 4. Exclusion, 5. Matura- tion, 6. Senescence)			
	IBP		Not used in assessment			
Forest continuity over time	(RB) Stand		Proportion of forest in 1956 (%)	Base year 1956		
	BP	Stand	Factor H. Areas with trees in the 1945 orthophoto and no signs of previous or subsequent agricul- tural use or soil disturbance as a consequence of reforestation.	Base year 1945		

4.1. NATIVE TREE SPECIES (IBP-RB)

Definition

Number of native tree species or genera present in the stand at any stage of development (including regeneration).

Sampling (differences and constraints) *RB*. All living tree species in the stand over 50 cm tall are recorded. The score recorded is the **number of distinct species** found in the entire stand.

IBP (Factor A). All living tree species in the stand over 50 cm tall are recorded **by genus**. The score recorded for the stand is the number of different **living or dead** genera found in one hectare. If two hectares are sampled, the average score is used.

Rationale

Maturity. In a natural forest it is likely that multiple tree species will coexist (Gosselin et al. 2004). In forests in more mature phases, more shade-tolerant companion species tend to appear in the vegetation strata below the canopy will gradually merge into the canopy. This slow merging process occurs as older trees lose part of their crown, leaving gaps that allow more light to enter, which is exploited by these species. Examples are species like *Sorbus* torminalis, S. domestica, Acer opalus, A. campestre, Tilia cordata, Prunus avium, Taxus baccata, etc.

Biodiversity potential. The biodiversity of the communities associated with trees depends on key structural differences such as how palatable their leaves are to insects and other phytophagous organisms, the hardness of the wood, the roughness and stability of the bark, the ability to form microhabitats, etc. While they vary from those in another genus in these and other ways, tree species within a genus will have similar associated communities of fauna and flora species. For example, the physicochemical characteristics of the bark of the genus Pinus determines the associated moss community, which is different from those associated, for example, with the genus Acer (Casas et al. 2003). Such specialisation is rare at tree species level. The same is true for birds. For example, most European tit species display a preference for either conifers or for broad-leaved trees, but do not distinguish between specific tree species (Camprodon 2013). Insects have a very diverse range of preferred habitats. For example, the larvae of certain Lepidoptera, Hymenoptera and beetles feed on genus-level nourishing plants, including arboreal genera. With deadwood, the associated saproxylic beetle community varies according to whether the deadwood is from coniferous or broad-leaved species.



Figure 5. Mediterranean mixed broad-leaved forest (oak and holm oak) and Aleppo pine (photo: Jordi Camprodon).

4.2. BASAL AREA (RB)

Definition

The average basal area (in m^2/ha) of all plots calculated taking all living trees of at least 17.5 cm DBH (trunk diameter measured at 1.30 cm above ground).

Sampling (differences and constraints) *RB*. All live trees at least 17.5 cm in diameter in each plot are sampled. The score recorded for the stand is the mean basal area of all sampled plots. IBP. Not sampled.

Rationale

Maturity. Basal area is a classic forest indicator, describing both the density of trees and their average size, providing a very simple indication of the standing biomass. The basal area will increase and decrease depending on the phase in the forest dynamics cycle, with lower values in the initial (gap, regeneration and occupation) and final phases (senescence phase) and higher values in the intermediate phases (exclusion and maturation).



Figure 6. The basal area of a stand increases significantly with the presence of mature trees (photo: Lluís Comas)

4.3. VERTICAL STRUCTURE (IBP-RB)

Definition

The number of vertical strata of vegetation present in each plot and at any stage of development, provided that in a given stratum the fraction of covered area is at least 20%.

Sampling (differences and constraints)

RB. Four vertical strata of equal height occupied by tree species are identified, plus a stratum of trees emerging from the dominant canopy. The score recorded for the stand is the mean of the strata of the sampled plots.

IBP (Factor B). The following strata are identified: herbaceous and semi-woody vegetation, very low (< 1.5 m), low (1.5-5 m), intermediate (5-15 m) and high (\geq 15 m) woody vegetation.

Rationale.

Maturity. A natural, mature forest will generally be irregular with various vertical strata (Bauhus et al. 2009). In the more mature stages of a forest, new strata tend to emerge because other, shade-tolerant, species appear and occupy the vegetation strata below the canopy. Over time, the canopies of the older trees in the upper canopy become less dense, allowing more light to enter the lower strata. Clearings may also open up in the canopy due to the death of a dominant tree, allowing the growth of new cohorts.

Biodiversity potential. A multi-stratified forest is vertically heterogeneous, favouring a wide range of species thanks to a diversity of microenvironments (with different sunlight, temperature and humidity levels) that can contain numerous taxonomic groups (lichens, mosses, aphyllophorous

fungi, birds, etc.). For example, there is a long-understood association between passerines and vertical vegetation structure (MacArthur and MacArthur 1961; Wilson 1974; Wiens 1989). Diversity increases if there is a well-developed bush and liana stratum, with the greatest diversity in holm oak and cork oak forests with cover of over 50% (Camprodon 2013). Similarly, a wealth of understory species leads to a greater diversity of insects associated with certain nutritious plants and of saprophytic, parasitic and mycorrhizal fungi. The shrub and liana stratum also provides shelter for ungulates and carnivores.. The forest bird community is usually very well associated with vertical stratification, for example, the tallest trees (more than 15 m), especially those that stand above those around them, facilitate the nesting of numerous birds of prey.



Figure 7. Coastal holm oak understory. The Mediterranean forest usually has a very abundant and diverse understory (photo: Lluís Comas).

4.4. DIAMETER CLASSES (RB)

Definition

This indicator is included in the RB protocol only and refers to the number of diameter classes (DC) of native tree species present in all the sampled plots.

Sampling (differences and constraints)

RB. The number of DC are counted from class 20, i.e. all living trees with DBH > 17.5 cm. The score recorded for the stand is the number of distinct DCs from all sampled plots.

IBP. Not sampled.

Rationale

Maturity. A natural forest usually has an irregular structure both in terms of diameters and tree height. A young forest typically has a smaller number of diameter classes with a reverse J-shaped, bimodal or fairly even distribution. As the forest grows, the number of classes increases and the proportion of trees in the smaller classes decreases. In the mature stages, the fall of a large tree allows a new cohort to grow, so the forest will contain trees in the smallest classes at the same time as very large trees.



Figure 8. Holm oak forest with various diameter classes in a LIFE BIORGEST stand (photo: Jordi Camprodon).

4.5. MEDIUM AND LARGE DEADWOOD (IBP-RB)

Definition

The quantity of standing or lying medium or large deadwood of any tree species found in the plot. The sampling method, indicators and constraints vary widely between the two protocols.

Sampling (differences and constraints) *RB*. The sampling threshold for deadwood is an DBH of at least 17.5 cm. No distinction is made between lying or standing deadwood. Two indicators are calculated from the data for each plot: the volume of deadwood (standing and lying) and the proportion of deadwood in relation to the volume of living trees. For both indicators the stand scale score is the maximum value for all sampled plots.

IBP. Separate counts are made of the number of dead standing trees or snags of at least 1 me-

ter in height (Factor C) or the number of lying dead trees (Factor D) of at least 1 meter in length (L). Medium deadwood (MDW) has an DBH of between 17.5 and 27.5 cm and large deadwood (LDW) has an DBH of at least 27.5 cm. Exceptionally, where the site quality is poor (type C) or for slow-growing species (genera *Arbutus, Acer, Pyrus, Sorbus*, etc.), deadwood with an DBH of 17.5 cm is treated as LDW.

Rationale

Maturity. Large deadwood is common in mature forests. It is an indicator of maturity because it is most abundant in the later phases of the forest dynamics cycle as larger trees, at the limit of their longevity, die off. The volume of deadwood as a proportion of the volume of living trees is greatest in the senescence phase. Deadwood, whether standing or lying, forms the basis of a complex food web allowing a succession of ecological processes, improving the integrity of the habitat and its natural balance, making it more resilient to external disturbances. Deadwood, whatever its size, can reduce erosion and is key for soil development, it stores carbon and water, it is an important source of energy and nutrients, helps certain species to germinate, and is an important habitat for decomposers and heterotrophic organisms (Harmon et al. 1986; Franklin et al. 1997; Kirby and Drake 1993; Samuelsson et al. 1994; McMinn and Crossley 1996; McComb and Lindenmayer 1999).

Biodiversity potential. Large deadwood is a key habitat for a wide range of saproxylic species (Müller and Bütler 2010). The dominant groups of saproxylic species include fungi, mosses, lichens, insects, amphibians, birds and mammals. A total of 25% of forest species depend on deadwood (Bobiec et al. 2005; Stokland et al. 2012), including three key ecological guilds for forest biodiversity: xylophages, detritivores and cavity-dwelling species. Of all the substrates, deadwood is probably the most critical for biodiversity (Jonsson and Siitonen 2013). Each organism plays a specific role in the decomposition cycle

of deadwood. Fungi successively break down sugars, cellulose and finally lignin. Some insects eat wood directly (xylophages), others consume fungi on deadwood, others are predators of the former, etc. More demanding species, or those with limited mobility, will only survive if there are substantial amounts of the required type of deadwood and it is well distributed throughout the stand (Bobiec et al. 2005). After fungi, saproxylic beetles are the most biodiverse species associated with deadwood. Oaks, for example, have been estimated to host about 900 species (Gilg 2012). Deadwood, especially large deadwood, also influences the diversity of epiphytic organisms: lichens and mosses (Hofmeister et al. 2015). For example, several species of epiphytic mosses have been identified as characteristic of advanced stages of wood decay in conditions where there are high levels of ambient humidity throughout much of the year (Crites and Dale 1998), for example, Buxbaumia viridis, B. aphylla and Calypogeia suecica. More mature forests contain greater volumes of deadwood, so the more naturalised the forest is, the richer in epiphytic mosses and lichens it will be (Boch et al. 2013; Ardelean et al. 2015). Epiphytes in turn form specific microhabitats for invertebrates. Their slow growth and limited dispersal capacity mean communities recover slowly from episodes of disturbance. Standing dead trees are important as a source of nesting holes for woodpeckers and of autogenically occurring cavities (raised bark, cracks in the trunk). The saproxylic community is the basis of a complex food chain. Invertebrates and birds prey on it and they in turn, together with parasites and parasitoids, regulate the populations of saproxylic organisms. For example, saproxylic fungi also form a microhabitat for saproxylic beetles that in their larval stage feed on the fruiting bodies of the fungi. In conclusion, a greater variety of deadwood in significant quantities leads to a greater diversity of species, a more complex network of interactions and more stable populations (Lachat et al. 2013).



Figure 9. Standing dead tree, retaining large branches. Saproxylic beetle exit holes and woodpecker feeding cavities can be seen (photo: Lluís Comas).

4.6. LARGE AND VERY LARGE TREES (IBP-RB)

Definition

The number of large or very large live trees present.

Sampling (differences and constraints)

RB. A tree is considered exceptional (very large) if its DBH (in cm) is greater than three times the dominant height (H_0 , in m) of the species in the plot. Example, if $H_0 = 15$ m, ED = 42.5 cm. The score, at stand scale, is the mean number of exceptional trees per hectare for all sample plots.

IBP (Factor E). A tree is considered large (LT) if the DBH is between 37.5 and 57.5 cm and very large (VLT) if its DBH is at least 57.5 cm. Exceptionally, where the site quality is poor (type C) or for slow-growing species (genera *Arbutus, Acer, Pyrus, Sorbus*, etc..), trees with an DBH of 37.5 cm are treated as VLT.

Rationale

Maturity. The number of exceptional trees is a good indicator of maturity because a tree takes a long time to reach an exceptional diameter, well over 100 years, usually more than 200 years. Live trees with exceptional diameters

make some of the most important contributions to the vertical structure of the forest. They are a vital refuge and resource for a rich variety of species and for ensuring communities continue to function. Exceptional trees are usually those that have reached the maximum possible height for a given site quality. As they age, these wide-crowned trees leave many open spaces that allow light to enter. This can be exploited by a wide range of shade-tolerant species that will occupy the intermediate strata.

Biodiversity potential. As a tree ages, it is more likely that a range of microhabitats will form, creating potential substrates for a great diversity of associated species, many of them saproxylic. Parts of the trunk and crown of large trees may die off, especially if they go into decline, but the living part can continue to grow for decades. While this is happening, new microhabitats will appear while others disappear. This dynamic results in a continuum of microhabitats (some very ephemeral) being maintained over very long- time scales, allowing for stable populations of a wide range of species, including rare or endangered species. Lichens and mosses are generally slow-growing organisms, some of them very slow, so species richness and abundance depend on maintaining the trees that serve as substrate and on stable microclimate conditions, for example, *Lobaria pulmonaria* large-thallus lichen that acts as a bioindicator of long-term stable conditions in forests (Gilg 2005).



Figure 10. Large holm oaks (Quercus ilex) are scarce, but older examples can become very large (photo: Lluís Comas).

4.7. TREE MICROHABITATS (IBP-RB)

Definition

The number of tree microhabitats (TreM) observed in living trees.

Sampling (differences and constraints)

RB. The **number of distinct TreMs from the 10 possible groups** (Annex A.1 and A.2). A TreM counts if it is found at least twice in the plots surveyed. If a tree has two different types of TreM, both are recorded; if the same tree has several TreMs of the same type, they are counted only once.

IBP (Factor F). The **number of live trees with TreM per hectare, provided they are different.** Each tree where a TreM is observed is classified under one of 15 possible groups (Annex A.1). All trees with TreMs observed are counted up to a maximum of two trees per hectare per TreM group. If a tree has different TreMs, each TreM type is counted; if the tree has several TreMs of the same type, it is counted once.

Rationale

Maturity. The abundance and diversity of tree microhabitats increase significantly with tree diameter and bark thickness and thus normally with tree age (Bütler and Lachat 2009; Vuidot

et al. 2011; Larrieu and Cabanettes 2012; Ellis 2012; Nascimbene et al. 2013; Larrieu et al. 2019). Consequently, live trees with TreM are usually large old trees associated with maturity. These trees contain different TreM that significantly increase the biodiversity of a wide range of species (especially invertebrates) and, therefore, promote and maintain certain ecological processes facilitates the resilience and natural balance of the habitat by establishing a complex network of interactions between species.

Biodiversity potential. Living trees, especially older ones, contain TreM that are essential for the survival of many species (Larrieu and Gonin 2008; Emberger et al. 2013). It is estimated that 20-40% of forest species in temperate and northern forests depend on or benefit from trees with TreM (Bobiec et al. 2005; Stokland et al. 2012; Bauhus et al. 2019), most notably saproxylic beetles (Parisi et al. 2019). Cavities are the microhabitats that host the most species of both invertebrates and vertebrates and those with more organic matter (Ranius 2002) are the richest in invertebrates. Mosses and lichens help conserve the ambient humidity of wood and soil, which is beneficial for other species such as fungi, vascular plants and invertebrates. They also provide habitats for small invertebrates such as nematodes and molluscs.



Figure 11. The 10 types of TreM listed in the Redbosques protocol and associated taxonomic groups (modified from Kraus et al. 2016, photos: Lluís Comas).

4.8. FOREST DYNAMICS (RB)

Definition

The presence of each of the six phases of the forest dynamics cycle throughout the entire stand. The forest dynamics cycle comprise six phases: 1. Gap, 2. Regeneration, 3. Occupation, 4. Exclusion, 5. Maturation, 6. Senescence.

Sampling (differences and constraints)

RB. The presence of a phase is recorded if it occupies an area of at least 200 m2, with the

exception of the regeneration phase which must be at least 100 m2. The score recorded for the stand is the sum of the values assigned to each phase: Clearing=2, Regeneration=1, Occupation=1, Exclusion=1, Maturation=2, Senescence=3.

IBP. Not sampled.

Rationale

Maturity. In the absence of major disturbances, in a forest with natural dynamics, every phase in the cycle will be observable. The structural

and ecological properties typical of mature forests appear gradually over time, resulting from the ecosystem's own dynamics, in a continuous cycle. Each generation of the dominant vegetation goes through successive structural stages, from new growth through to the complete renewal of the canopy once all the individuals from the initial generation die off. Different ecological processes take place in the tree ecosystem at each phase of the cycle. The last stages (maturity and senescence) are the most important in terms of forest maturity as they require up to hundreds of years. Bauhus et al. (2009) estimate that in forests managed for timber 10-40% of the cycle does not occur, i.e., they are kept in the early stages of the cycle.



Figure 12. Large gap in a holm oak stand with regeneration of yew, holm oak and different herbaceous species (photo: Jordi Camprodon).

4.9. FLOWER-RICH OPEN AREAS (IBP)

Definition

The proportion of the surface area containing open spaces with flowering vegetation (forest clearings, sparse forest, open spaces on the forest edge). These may be permanent or temporary, natural or due to management.

Sampling (differences and constraints) *RB*. Not sampled.

IBP (Factor G). Record the surface area of clearings and areas of sparse vegetation in the entire stand. A score is given if the area occupied by flowering species is between 1% and 5%.

Rationale

Biodiversity potential. Forest biodiversity requires a certain proportion, albeit low, of open spaces that allow flowering species to be relatively permanently present (monotones, corridors, etc.). Many

forests and saproxylic fauna need open, sunny spaces at some point in their life cycle. Some species of saproxylic beetles feed on flower nectar and pollen in their adult phase, for example. At the stand scale, there needs to be enough open space to maintain viable populations of these species, but not so much that it would compromise the light, temperature and humidity levels that typify dense forests, thereby endangering the associated biodiversity.



Figure 13. Open spaces at the forest edge, colonised by flowering species (photo: Lluís Comas).

5. CONTEXT INDICATORS

5.1. FOREST CONTINUITY (IBP-RB)

Definition

The stand is deemed to be old-growth forest if the land was already tree covered in the mid-20th century and its use has not changed since then.

Sampling (differences and constraints)

RB. The percentage of the land covered by trees in 1956-57 per the orthophotomap developed from aerial photographs taken in that year is recorded. This indicator is used to assess the historical human footprint together with the agricultural, livestock and forestry uses of the land.

IBP (Factor H). The stand is deemed to be old-growth forest if the land was tree-covered per the 1945 orthophoto and there are no indications of previous or subsequent agricultural use or soil disturbance as a result of reforestation. For this indicator, it is also necessary to note in the field if there are evident signs of forest discontinuity (walls, terraces) in all or part of the stand, or evidence of forest continuity in stands that were clear of trees in 1945 (very old trees retained on the edges of former pastures, rocky areas where trees were not removed, etc.); soil disturbance throughout entire reforested areas (subsoiling, ploughing between rows, the uprooting of vines); any historical document that indicates the age of the forest.

Rationale

Maturity. The maturity of a forest is closely linked to the state of the soil. Forest soil takes many decades to mature and stabilise. Any disturbance of old soil has immediate consequences, and it can take many decades, even centuries, to recover.

Biodiversity potential. Some species of flora grow only in forests and require forest conti-

nuity because they have a very low dispersal ability or a low capacity to adapt to non-forest soils. These species will not, therefore, be found in forests that are recently established on pastures or abandoned agricultural land (Hermy et al. 1999; Hermy and Verheyen 2007; Dupouey et al. 2002a and 2002b).

Figure 14. Forest continuity over more than 60 years. Comparison between aerial photographs taken in 1956 and the present day (source: Institut Cartogràfic i Geològic de Catalunya).

5.2. AQUATIC HABITATS (IBP)

Definition

The presence of different types of aquatic habitat in the stand or its immediate vicinity.

Sampling (differences and constraints) *RB*. Not sampled.

IBP (Factor I). The presence of the different types listed in Annex A.3 is recorded (the maximum score is given if there are at least two different types).

Rationale

Biodiversity potential. Freshwater aquatic systems

interconnect and interact ecologically with the ecosystems through which they flow and act as biological connectors (Gregory et al. 1991; Wohl 2016). Their physical, chemical and biological characteristics depend on and reflect the state of the ecosystems of the basin as a whole. At the stand scale, shade cast by riparian trees regulates water temperature, limiting algal blooms, slowing decomposition processes and eutrophication, and maintaining suitable environmental conditions for amphibian and fish. Forest bats drink and hunt by flying over bodies of water. Many species of birds, mammals and reptiles are semi-aquatic, such as the white-footed shrew, the desman, the water rat, the otter and the water snake. Duck and heron roost and breed in the trunks or crowns of riverside trees. Amphibians need watercourses and watering holes for breeding, and dense forest near the streams where they live to keep these habitats sufficiently cool and damp (especially in summer).

Figure. 15. Breeding colony of herons with up to 300 nests in an alder grove on the banks of the Ter River (photo: Jordi Bas).

5.3. ROCKY HABITATS (IBP)

Definition

The presence of different types of rocky habitat in the stand or its immediate vicinity.

Sampling (thresholds and particularities) *RB*. Not sampled.

IBP (Factor J). The presence of the different types listed in Annex A.4 is recorded (the maximum score is given if there are at least two different types), provided the habitat accounts for a total area of at least 20 m²/ha.

Rationale

Biodiversity potential. Forest biodiversity may depend in part on rocky. Rocky habitats have distinctive characteristics (type of rock, humidity, thermal inertia) providing a suitable

environment for different forest species: a substrate for mosses and lichens, micro-soils for rock-growing flora, shade, shelter, refuge for numerous reptiles, amphibians or arthropods.

Figure 16. Top: dry wall in a Life BIORGEST holm oak stand. Bottom: rocky outcrop in the shade of a holm oak forest with Ramonda myconi (photos: Jordi Camprodon).

6. THRESHOLDS

he threshold values used to score each indicator are set out separately for each protocol (Table 2). Four levels are proposed for each indicator: very low, low, medium and high. The threshold values in the table are based on those detailed in the original Redbosques (EUROPARC-Spain 2020b) and IBP (Baiges et al. 2022) protocols.

The threshold applied for structural attributes depends, to a large extent, on the taxonomic group or processes that need to be protected (Bauhus et al. 2009). In the absence of more detailed information, the solution is to maintain the features and attributes needed to support as many processes as possible at the same time, thereby providing habitats for a wide range of species. This is the philosophy behind the IBP. However, for most forest ecosystems, our understanding remains guite limited of the guantity of these features needed, and how they need to be distributed in space and time, in order to meet certain biodiversity and maturity targets (Bauhus et al. 2009; Müller and Bütler 2010; Bouget et al. 2013; Larrieu et al. 2019). This is especially true in the Mediterranean region. For example, since habitat requirements differ between species and for different types of forest, it is almost impossible to identify deadwood thresholds that guarantee the survival of the entire saproxylic species community (Ranius and Jonsson 2007; Jonsson and Siitonen 2013). In addition, maintaining stable communities of saproxylic species depends not only on the quantity but also on the quality of the structural features, and on a forest structure that ensures these features are also maintained over time.

Figure 17. Girdled and felled deadwood to encourage biodiversity (photos: Jordi Camprodon).

In practice, it is easier to define a threshold when it is based on a single species. However, it is much more useful to try and set thresholds at the community level. It makes sense, therefore, to consider as many species as possible when setting habitat thresholds for maintaining the entire community of species that depend on, for example, deadwood (Lachat et al. 2013; Bouget et al. 2013).

Table 2 shows the proposed thresholds for both protocols: IBP thresholds for biodiversity carrying capacity indicators and RB thresholds for stand maturity indicators. For IBP, the minimum values would be those considered compatible with multifunctional forest management: IBP thresholds corresponding to a score of 5 for each indicator. It should be noted that, although achieving this minimum score can produce a qualitative leap in terms of biodiversity for many taxonomic groups, it does not mean that increasing, for example, the quantity of wood would further boost the associated biodiversity.

For maturity, the proposed thresholds are those associated with natural dynamics, i.e., greater maturity and a smaller human footprint, in short, management aimed at conserving the functions and processes inherent to natural dynamics and the associated biodiversity. The thresholds for each indicator are based on a range established in the original protocol (EUROPARC-Spain. 2020b), with scores from 0 to 10, as part of the work of the LIFE Redbosques project. For example, the range of values established for the volume of deadwood for holm oak forests is between 5 and 25 m³/ha, i.e., below 5 m³/ha the indicator score for deadwood is very low, above 25 m³/ha the score is high, while between 5 and 25 the score is low to medium. A forest is considered mature if the reference scores are "high".

TABLE 2.

Threshold values for each indicator for biodiversity hosting capacity and maturity. For details of each variable see Table 1.

			Threshol	d score	
		Very low	Low	Medium	High
Capacity to hose Maturity	t biodiversity	0 0-0.9	1 1.0-3.9	2 4.0-6.9	5 7.0-10
Indicator	Variable				
Native tree species	BP No. of genera (living or dead individuals) BP No. of live species	0-1 <3	2 3-4	3-4 5-6	≥5 ≥7
Basal area	🔞 Basal area (living trees) (m²/ha)	<21.5	21.5-25.9	26-30.4	≥30.5
Diameter classes	RB No. of classes ⁽¹⁾	<6 <8 <6	6-8 8-10 6-8	9-11 11-13 9-11	≥11 ≥13 ≥11
Vertical structure	BP No. of vegetation strata BB No. of tree strata	<2 <2.2	2 2.2-2.8	3-4 2.9-3.4	5 ≥3.4
Large and very large trees	BP Number of large (LT) and very large (VLT) trees (trees/ha)	<1 VLT and LT	<1 VLT and ≥1 LT	1-4 VLT	≥5 VLT
	B Number of exceptional trees (trees/ha) ⁽¹⁾	<14 <23 <33	14-25.9 23-31.9 33-41.9	26-37.9 32-40.9 42-50.9	≥38 ≥41 ≥51
Medium and large deadwood	B Number off medium (MDW) and large (LDW) standing (trees/ha)	<1 LDW and MDW	<1 LDW and ≥1 MDW	1-2 LDW	≥3 LDW
	Number of medium (MDW) and large (LDW) lying (pieces/ha)	<1 LDW and MDW	< LDW and ≥1 MDW	1-2 LDW	≥3 LDW
	B Number standing and lying (m ³ /ha) ⁽¹⁾	<14 <14 <8	14-25.9 14-25.9 8-16.9	26-37.9 26-37.9 17-25.9	≥38 ≥38 ≥26
	RB Deadwood as % of live trees	<7.5	7.5-14.9	15-22.4	≥22.5
Tree microhabitats	Irees with TreMs (tree/ha)Number of different types	<2 <4	2 4	3-7 5-6	≥8 ≥7
Flower-rich open areas	BP Proportion of area without tree cover (%)	0	0	<1 o >5	1-5
Dynamic	Silvogenetic phases (phase) ⁽²⁾	1 and/or 2	3 and/or 4	5 and 6	All
Forest continuity over time	BP Forest before 1945 (value) BP Proportion of forest before 1956 (%)	0 0-10	0 11-25	2 26-75	5 ≥76

⁽¹⁾ Score by habitat and by order: 42.84 - Aleppo pine forests (Pinus halepensis); 41.7&1 - Oak groves (Quercus humilis) or hybrids; 45.3 - Evergreen oak and holm oak (Quercus ilex or Q. rotundifolia)

⁽²⁾ Six distinct silvogenetic phases are observed: 1) gap, 2) regeneration, 3) occupation, 4) exclusion, 5) maturity and 6) senescence

7. COMBINED ASSESSMENT FIELD PROTOCOL

o changes are made to the respective field protocols for each assessment system, except for some details described in this section. For assessing maturity, the plot sampling system should be used (LIFE Redbosques 2018). This document describes the field methodology and the calculations to be made to obtain the maturity indicators. In order for the IBP assessment to be compatible with the Redbosques assessment, the IBP sampling system for plots must also be followed (*https://cpf.gencat.cat/ca/cpf_03_linies_actuacio/cpf_transferencia_coneixement/Index-Biodiversitat-Potencial/ documents-i-publicacions-relacionades-amb-libp/ Fitxes_i_protocol_IBP/*).

The field table used for the combined sampling methodology is contained in Annex A.5. The data sheet is divided into two sections. The first is for stand scale data, comprising the sum of the indicators collected at plot scale as well as those collected only at stand scale, such as the IBP context factors (Factors H, I and J) or, for example, the RB indicator for the number of different diameter classes. In the plot section, the size of the plots must be decided in order to determine the number of plots needed to sample the required area.

The area sampled must cover at least one hectare and represent between 15% and 50% of the total area, i.e., for every 6 hectares, approximately, at least one hectare must be sampled. It is highly recommended to carry out the sampling with circular plots, therefore, if plots of 25.2 metres radius are used, five plots are needed, if the plots are of 32.6 m radius, three plots are needed, and so on. It is recommended that no fewer than three plots per stand be used.

For IBP sampling, there is no upper limit for factors C and D (standing and lying deadwood) and factor E (very large trees), even if the threshold for obtaining the maximum score of 5 has been reached. The only upper limit applied is for factor F (TreM), if the threshold value of two trees with the same microhabitat is reached. For the remaining factors, A, H, I and J, the original sampling methodology should be followed.

With this sampling approach, the complete assessment takes more time than that proposed in the original versions of the sampling protocol.

The differences in the sampling approach with respect to the respective original protocols are:

- In each plot, count the number of live trees by diameter class (DC) and species starting at DC 20 (ND>17.5 cm). For DC 20 and DC 25, count only the trees up to the 10 m radius.
 From DC 30 to DC 55, all trees within the chosen sampling radius (25.2 m if five plots are sampled, 32.6 m if three plots are sampled, etc.) must be counted. From DC 60 onwards, the DBH must be measured and noted. The measurements can be taken with a tree caliper or forestry tape measure.
- Obtain the dominant height (H₀) in each plot and for each species (normally one species, or two if the CCF of the second is at least 30%). H₀ is calculated from the average of the three thickest trees in the plot. This figure, with the number of trees per DC and per species, is used to calculate the volume, including bark, of live trees. This can then be used to obtain the deadwood to live wood ratio.
- For lying deadwood, of all pieces with a diameter of at least 17.5 cm, measure the length up to this diameter and the diameter of the trunk at half this length. For the standing dead trees present in the plot, measure the

normal diameter and height of the trunk. This information can be combined with the total number of pieces of lying and standing deadwood to calculate the total deadwood volume.

 For the IBP indicator for live trees with TreM, the number of trees is recorded by TreM type observed, up to a maximum of two trees/ha per TreM group, based on the 15 types listed in Annex A.1 and A.2. If a tree has different TreMs, each TreM type is counted as one tree; if the tree has several TreMs of the same type, it is counted once. For Redbosques, record the number of different TreM detected in all the plots, based on the 10 types detailed in Annex A1 and A.2. A TreM type counts if there are at least two per hectare. If a tree has two different types of microhabitats, it will be recorded twice.

Figure 18. IBP and RedBosques protocol field sampling (photo: Lluís Comas)

8. REFERENCES

Ardelean, I. V., Keller, C., Scheidegger, C. 2015. Effects of management on lichen species richness, ecological traits and community structure in the Rodnei Mountains National Park (Romania). PLoS One, 10 (12): e0145808.

Avila-Cabadilla, L. D., Stoner, K. E., Henry, M., y Añorve, M. Y. A. 2009. Composition, structure and diversity of phyllostomid bat assemblages in different successional stages of a tropical dry forest. Forest Ecol. Manage., 258 (6): 986–996. https://doi.org/10.1016/j. foreco.2008.12.011

Baiges, T., Cervera, T., Palero, N., Gonin, P., Larrieu, L. 2022. El Índice de Biodiversidad Potencial (IBP) como herramienta de apoyo a la gestión forestal: fundamentos y aplicaciones en Cataluña. Actas del 8ª CFE. SECF.

Bauhus, J., Puettmann, K., Messier, C. 2009. Silviculture for old-growth attributes. Forest Ecol. Manage., 258(4), 525–537. https://doi.org/ 10.1016/j.foreco.2009.01.053

Bauhus, J., Forrester, D. I., Gardiner, B., Jactel, H., Vallejo, R., Pretzsch, H. 2017. Ecological Stability of Mixed-Species Forests. In: Pretzsch, H., Forrester, D., Bauhus, J. (eds). Mixed-Species Forests. Springer, Berlin, Heidelberg. *https://doi. org/10.1007/978-3-662-54553-9_7*

Bauhus, J., Baber, K. and Müller, J. 2019. Dead Wood in Forest Ecosystems. Oxford Bibliographies. Ecology. Oxford Bibliographies. *https:// doi.org/10.1093/OBO/9780199830060-0196*

Bobiec, A., Gutowski, J. M., Laudenslayer, W. F., Pawlaczyk, P., Zub, K. 2005. The afterlife of the tree WWF Poland, Warszawa. Boch, S., Prati, D.m Hessenmöller, D., Schulze, E., Fischer, M. 2013. Richness of lichen species, especially of threatened ones, is promoted by management methods furthering stand continuity. PLoS One, 8(1): e55461.

Bouget, C., Larrieu, L., Nusillard, B., Parmain, G., 2013. In search of the best local habitat drivers for saproxylic beetle diversity in temperate deciduous forests. Biodivers. Conserv., 22: 2111–2130. https://doi.org/10.1007/s10531-013-0531-3

Bütler, R., Lachat, T. 2009. Wälder ohne Bewirtschaftung: eine Chance für die saproxylische Biodiversität. Schweizerische Zeitschrift für Forstwesen 160 (11): 324–333. *https://doi. org/10.3188/szf.2009.0324*

Camprodon J., 2013. Ecologia i conservació dels ocells forestals. Un manual de gestió de la biodiversitat en boscos catalans. CTFC & Departament d'Agricultura, Ramaderia, Pesca, Alimentació i Medi Natural de la Generalitat de Catalunya. 223 pp.

Casas C., Brugués M., Cros R. M. 2003. Flora dels briòfits dels Països Catalans. Vol. I, molses. Institut d'Estudis Catalans, Barcelona.

Crites, S., Dale, M. R. T. 1998. Diversity and abundance of bryophytes, lichens, and fungi in relation to woody substrate and successional stage in aspen mixedwood boreal forests. Can. J. Bot., 76: 641-651.

De la Peña-Cuéllar, E., Stoner, K. E., Avila-Cabadilla, L. D., Martínez- Ramos, M., Estrada, A. 2012. Phyllostomid bat assemblages in different successional stages of tropical rain forest in Chiapas, Mexico. Biodiversity and Conservation, 21(6): 1381–1397. https://doi.org/10.1007/s10531-012-0249-7 Dupouey, J. L., Dambrine, E., Laffite, J. D., Moares, C., 2002a. Irreversible Impact of Past Land Use On Forest Soils and Biodiversity. Ecology, 83: 2978–2984. https://doi.org/10.1890/0012-9658(2002)083[2978:I-IOPLU]2.0.CO;2

Dupouey, J.-L., Sciama, D., Koerner, W., Dambrine, E. 2002b. La végétation des forêts anciennes. Revue forestière française, 54 (6): 521-532.

Ellis, C. J. 2012. Lichen epiphyte diversity: A species, community and trait-based review. Perspectives in Plant Ecology, Evolution and Systematics, 14: 131–152. *https://doi.org/10.1016/j.ppees.2011.10.001*

Emberger, C., Larrieu, L., Gonin, P., 2013. Dix facteurs clés pour la diversité des espèces en forêt. Comprendre l'Indice de Biodiversité Potentielle (IBP). Document technique. Paris. Institut pour le développement forestier. 56 pp.

EUROPARC-España. 2020a. Bosques maduros mediterráneos: características y criterios de gestión en áreas protegidas Ed. Fundación Fernando González Bernáldez, Madrid. *https://red-bosques.eu/documentacion-tecnica/*

EUROPARC-España. 2020b. Red de Rodales de Referencia. Manual técnico. Ed. Fundación Fernando González Bernáldez, Madrid. *https://redbosques.eu/documentacion-tecnica/*

FAO. 2020. Global Forest Resources Assessment 2020 – Key findings. Rome. *https://doi.org/10.4060/ca8753en*

Ferris, R., Humphrey, J. W. 1999. A review of potential biodiversity indicators for application in British forests. Forestry, 72: 313–328. *https://doi. org/10.1093/forestry/72.4.313*

Franklin, J. F., Berg, D. R., Thornburgh, D. A., Tappeiner, J. C. 1997. Alternative silvicultural approaches to timber harvesting: Variable retention systems. In Kohm K. A., Franklin J. F. (eds.). Creating a Forestry for the 21st Century: The Science of Forest Management. Island Press, pp. 111–139. Gao, T., Nielsen, A. B., Hedblom, M. 2015. Reviewing the strength of evidence of biodiversity indicators for forest ecosystems in Europe. Ecological Indicators, 57: 420–434. *https://doi. org/10.1016/j.ecolind.2015.05.028*

Gilg, O. 2005. Old-Growth Forests: characteristics, conservation and monitoring. L'Atelier technique des espaces naturels & Réserves Naturelles de France. 52 pp.

Gonin, P., Larrieu, L., Martel, S. 2012. L'Indice de Biodiversité Potentielle (IBP) en région méditerranéenne. Forêt Méditerranéenne, XXXIII n°2, pp. 133-141.

Gosselin, F., Nageleisen, L-M., Bouget, C., 2004. Réflexions pour mieux gérer le bois mort en faveur de la biodiversité. Forêt entreprise, 438: 26-29.

Gregory S.V., Swanson F.J., McKee W.A., Cummins, K.W. 1991. An ecosystem perspective of riparian zones. BioScience, 41: 540-551. 10.2307/1311607.

Gustafsson, L., Bauhus, J., Asbeck, T., Augustynczik, A. L. D., Basile, M., Frey, J., Gutzat, F., Hanewinkel, M., Helbach, J., Jonker, M., Knuff, A., Messier, C., Penner, J., Pyttel, P., Reif, A., Storch, F., Winiger, N., Winkel, G., Yousefpour, R., Storch, I. 2019. Retention as an integrated biodiversity conservation approach for continuous-cover forestry in Europe. Ambio. *https://doi.org/10.1007/s13280-019-01190-1*

Harmon, M. E., Franklin, J. F., Swanson, F. J., Sollins, P., Gregory, S.V., Lattin, J. D., Anderson, N. H., Cline, S. P., Aumen, N. G., Sedell, J. R., Lienkaemper, G. W., Cromack, K., Cummins, K. W. 1986. Ecology of coarse woody debris in temperate ecosystems. Advances in Ecological Research, 15: 133–302. https://doi.org/10.1016/ S0065-2504(08)60121-X Hermy, M., Honnay, O., Firbank, L., Grashof-Bokdam, C., Lawesson, J. E. 1999. An ecological comparison between ancient and other forest plant species of Europe, and the implications for forest conservation. Biological Conservation, 91: 9-22. *https://doi.org/10.1016/S0006-3207(99)00045-2*

Hilmers, T., Friess, N., Bässler, C., Heurich, M., Brandl, R., Pretzsch, H., Seidl, R., Müller, J. 2018. Biodiversity along temperate forest succession. Journal of Applied Ecology, 55(6): 2756–2766. https://doi.org/10.1111/1365-2664.13238

Hofmeister, J., Hosek, J., Brabec, M., Dvorák, D., Beran, M., Deckerová, H., Burel, J., Koiž, M., Borovioka, J., Boták, J., Vašutová, M., Malíoek, J., Palice, Z., Syrovátková, L., Steinová, J., Cernajová, I., Holá, E., Novozámská, E., Cížek, L., Iarema, V., Baltaziuk, K., Svoboda, T. 2015. Value of old forest attributes related to cryptogam species richness in temperate forests: A quantitative assessment. Ecological Indicators, 57: 497-504. https://doi.org/10.1016/j.ecolind.2015.05.015

Jacobs, J. M., Spence, J. R., Langor, D. W. 2007. Influence of boreal forest succession and dead wood qualities on saproxylic beetles. Agricultural and Forest Entomology, 9 (1): 3–16. *https://doi. org/10.1111/j.1461-9563.2006.00310.x*

Jonsson, G., Siitonen, J. 2013. Managing for target species. in: Kraus D., Krumm F. (eds). Integrative approaches as an opportunity for the conservation of forest biodiversity. European Forest Institute. 284 pp.

Kirby, K. J., Drake, C. M., 1993. Dead Wood Matter: The Ecology and Conservation of Saproxylic Invertebrates in Britain. English Nature Science, No. 7. Peterbourough, UK.

Kraus D., Bütler R., Krumm F. Lachat T. Larrieu L. et al. 2016. Catalogue of tree microhábitats –Field reference list. Integrate + Technical Paper Nr. 13. 16Ss. Kuusinen, M., Siitonen, J. 1998. Epiphytic lichen diversity in old- growth and managed Picea abies stands in southern Finland. Journal of Vegetation Science, 9(2): 283–292. *https://doi.org/10.2307/3237127*

Lachat T, Bouget C, Bütler, R., Müller J. 2013. Deadwood: quantitative and qualitative requirements for the conservation of saproxylic biodiversity. in: Kraus D., Krumm F. (eds). Integrative approaches as an opportunity for the conservation of forest biodiversity. European Forest Institute. 284 pp.

Larrieu, L., Gonin, P., 2008. L'indice de biodiversité potentielle (IBP): une méthode simple et rapide pour évaluer la biodiversité potentielle des peuplements forestiers. Revue Forestière Française, 6: 727-748. https://doi.org/ 10.4267/2042/28373ff.ffhal-03449570f

Larrieu, L., Cabanettes, A. 2012. Species, live status, and diameter are important tree features for diversity and abundance of tree microhabitats in subnatural montane beech-fir forests. Canadian Journal of Forest Research, 42 (8): 1433–1445. *https://doi.org/10.1139/x2012-077*

Larrieu, L., Gosselin, F., Archaux, F., Chevalier, R., Corriol, G., Dauffy-Richard, E., Deconchat, M., Gosselin, M., Ladet, S., Savoie, J. M., Tillon, L., Bouget, C. 2019. Assessing the potential of routine stand variables from multi-taxon data as habitat surrogates in European temperate forests. Ecological Indicators, 104: 116–126. https://doi.org/10.1016/j.ecolind.2019.04.085

LIFE Redbosques. 2018. Manual de campo para la identificación de rodales de referencia. Fase II: Identificación mediante parcelas. Ed. Fundación González Bernáldez, Madrid. Proyecto LIFE Redcapacita_2015. Deliverable B3.2. 53 pp. https:// redbosques.eu/documentacion-tecnica/ Lindenmayer, D. B., Margules, C. R., Botkin, D. B. 2000. Indicators of biodiversity for ecologically sustainable forest management. Conserv. Biol., 14 (4): 941–950. *https://doi.org/10.1046/j.1523-1739.2000.98533.x*

Lindenmayer, D. B., Franklin, J. F., Fischer, J. 2006. General management principles and a checklist of strategies to guide forest biodiversity conservation. Biol. Conserv., 131 (3): 433–445. *https:// doi.org/10.1016/j.biocon.2006.02.019*

Macarthur, R. H., Macarthur, R. W. 1961. On bird species diversity. Ecology, 42: 594-598.

McComb, W., Lindenmayer, D. 1999. Dying, dead, and down trees. In: Hunter Jr, M. L. (ed.). Maintaining biodiversity in forest ecosystems. Cambridge, Cambridge University Press, pp. 335-372.

McMinn, J. W., Crossley, D. A. 1996. Biodiversity and Coarse Woody Debris in Southern Forests. USDA Forest Service General Technical Report SE-94, 156 pp.

Müller, J. Bütler, R. 2010. A review of habitat thresholds for dead wood: a baseline for management recommendations in European forests. European Journal of Forest Research 129:981–992. https://doi.org/10.1007/s10342-010-0400-5

Nascimbene, J., Ylisirniö, A-L, Pykälä, J., Giordani P. 2013. Lichens: sensitive indicators of changes in the forest environment. In: Kraus D., Krumm F. (eds). Integrative approaches as an opportunity for the conservation of forest biodiversity. European Forest Institute. 284 pp.

Parisi, F., Di Febbraro M., Lombardi, F., Biscaccianti, A. B., Campanaro, A., Tognetti, R., Marchetti, M. 2019. Relationships between stand structural attributes and saproxylic beetle abundance in a Mediterranean broadleaved mixed forest. Forest Ecol. Manage., 432: 957-966. https://doi. org/10.1016/j.foreco.2018.10.040 Ranius T. 2002. Influence of stand size and quality of tree hollows on saproxylic beetles in Sweden. Biol. Conserv., 103: 85–91. *https://doi.org/10.1016/S0006-3207(01)00124-0*

Ranius, T., Jonsson, M. 2007. Theoretical expectations for thresholds in the relationship between number of wood-living species and amount of coarse woody debris: A study case in spruce forests. Journal for Nature Conservation, 15: 120–130. *https://doi.org/10.1016/j.jnc.2007.02.001*

Redecker, D., Szaro, T. M., Bowman, R. J., Bruns, T. D. 2001. Small genets of Lactarius xanthogalactus, Russula cremoricolor and Amanita francheti in late-stage ectomycorrhizal successions. Molecular Ecology, 10(4): 1025–1034. *https:// doi.org/10.1046/j.1365-294X.2001.01230.x*

Rossi, M., Vallauri, D. 2013. Évaluer la naturalité. Guide pratique version 1.2. Rapport WWF France, 154 pp.

Samuelsson, J., Gustafsson, L., Ingelog, T. 1994. Dying and Dead Trees: A Review of their Importance for Biodiversity. Swedish Environmental Protection Agency Report 4306, Uppsala.

Schowalter, T. D. 1995. Canopy arthropod communities in relation to forest age and alternative harvest practices in western Oregon. Forest Ecol. Manage., 78: 115–125. *https://doi. org/10.1016/0378-1127(95)03592-4*

Stein, A., Kreft, H. 2015. Terminology and quantification of environmental heterogeneity in species-richness research. Biological Reviews, 90 (3): 815–836. https://doi.org/10.1111/brv.12135 Stokland, J., Siitonen, J., Jonsson, B. G. 2012. Biodiversity in dead wood. Cambridge University Press. 509 pp.

Vuidot, A., Paillet, Y., Archaux, F., Gosselin, F. 2011. Influence of tree characteristics and forest management on tree microhabitats. Biological Conservation, 144 (1): 441–450. https://doi.org/ 10.1016/j.biocon.2010.09.030

WCFSD, 1999. Summary report: World Commission on Forests and Sustainable Development, World Commission on Forests and Sustainable Development, Winipeg, Canada. 40 pp.

Wiens, J. A. 1989. The ecology of bird communities. Volume 1 i 2. Cambridge University Press, Cambridge. 539 pp.

Wilson, M. F. 1974. Avian community organization and habitat structure. Ecology, 55: 1017-1029.

Wirth C., Gleixner G., Heimann M. (eds.). 2009. Old-Growth Forests: Function, Fate and Value. Springer-Verlag. Berlin. Ecological Studies, Vol. 207.

Wohl, E. 2016. Messy rivers are healthy rivers: The role of physical complexity in sustaining ecosystem processes. In Constantinescu, G., Garcia, M., Hanes, D. (eds.). River Flow. CRC Press. Iowa City, pp. 24–27.

9. ANNEXES

A.1. TREE MICROHABITATS

THE THREE TYPOLOGIES OF THE INDEX OF BIODIVERSITY POTENTIAL (IBP) USED IN V3.0

A.2. COMPARISON OF THE TWO TREE MICROHABITAT CLASSIFICATION SYSTEMS (15 TYPES VS. 10 TYPES)

Classification of 15 types (based on Larrieu et al. 2018)	Classification of 10 types (based on Kraus et al. 2016)
1. Woodpecker breeding cavities	CP. Woodpecker breeding cavities
2. Rot-holes	OC: Other cavities
3. Insect galleries and bore holes	
4. Concavities	
5. Exposed sapwood only	CO: Bark
6. Exposed sapwood and heartwood	DH: Injuries and wounds
7. Crown deadwood	MM: Deadwood
8. Burrs and cankers	FC: Deformation / growth form
9. Twig tangles	
10. Perennial fungal fruiting bodies	HO: Fungal bodies
11. Ephemeral fungal fruiting bodies	
12. Epiphytic or parasitic crypto- and phanerogams	EP: Epiphytes
13. Nests	NI: Nests
14. Microsoils	OT: Others
15 Fresh exudates	

A.3. AQUATIC HABITATS

	TYPOLOGY OF AQU	ATIC HABITATS
Types of aquatic habitats	How to identify them?	Comments
Spring or seepage	A point where water emerges from underground. These habitats are limited to the point where the water seeps through. They may be a spring or an area of diffuse seepage on a slope or among rocks. Springs and groundwater seepage can give rise to small streams or marshy areas, which are defined as other habitat types).	At the point where the spring emerges, the water has the same characteristics as its underground water table: usually cool, even in summer, with a high oxygen concentration, although hot springs also exist. In either case, its characteristics differ from those of the streams or marshy areas fed by the spring. The biodiversity of these habitats is quite specific, with some species of high heritage value, such as the Killarney fern (<i>Trichomanes speciosum</i>) or the Willow-leaved loosestrife (<i>Lysimachia ephemerum</i>).
Small stream, unmaintained humid ditch or small channel	This type includes: > Small natural streams (width < 1 m) in the uppermost reaches of the hydrographic network. > Small man-made streams (width < 1 m), especially drainage or irrigation channels.	River Rivulet Secondary brancifes main channes
Stream	No more than 1 to 8 m in width, streams are just downstream from small streams in the hydrographic network. As their catchment basins are small, their flow rate is low.	The morphology and characteristics vary and are closely dependent on the rate of water flow, which in turn depends on the slope. They range from mountain torrents and cascades to meanders in lowland reaches. Rivers are mainly permanent, although some can flow intermittently in the Mediterranean region, especially the secondary river branches. Water flow may be either permanent or intermittent in small streams. The presence of specific and often hydrophilic vegetation on riverbanks or riverbeds indicates a habitat which is under water for part of the year. Water can therefore be present whether floods have occurred or not. Riverbanks and riverbeds may be governed by natural dynamics or modified by human activities. However, rivers with consolidated banks and regularly maintained ditches host far fewer species (a rough natural substrate providing shelters and anchorage is especially important when there
River, estuary or delta	Over 8 m in width, these are downstream from streams. They may be subdivided into a main channel and secondary branches connected to it. Rivers may be tributaries of other rivers, while major rivers flow into the ocean or sea, sometimes forming an estuary (in which the tides mix freshwater and seawater) or a delta (when a river divides into several channels near its outlet due to the accumulation of sediment).	Streams can the narrow enough to flow entirely beneath forest vegetation. Any stream or river has alternating stretches of faster or slower-moving water of different heights. As long as the water is visibly flowing, these successive stretches cannot be individualised. However, stretches with stagnant water trapped in hollows are considered as "shallow ponds and water bodies" when the water column is shallow (1 to 3 m on average) and as" lakes and deep water bodies" when the water is deeper.
Backwater	A section of a river corresponding to a former channel, which is usually disconnected from the main or secondary riverbed, except very occasionally when the river is in spate (see diagram below).	High seasonal variations in the volume of water and its characteristics (temperature, etc.), which influence the vegetation and fauna present.

	TYPOLOGY OF AQU	
Types of aquatic habitats	How to identify them?	Comments
Lake or deep water body	A large, deep inland water body. (Natural or man-made).	At depths of 15 m or more, sunlight no longer penetrates and the water temperature rapidly decreases. Plant species cannot develop below 15 m. The water layers can be mixed by seasonal effects. In the case of artificial lakes (gravel pits, abandoned quarries, reservoirs, dams for irrigation, low water replenishment, hydropower, etc.), the materials used to build up the banks considerably reduce the number of species present. They are still used by wildlife for some
Pond, lagoon or shallow water body	Although they are shallow (1 to 3 m in depth on average), the warming action of the sun does not always reach to the bottom of these water bodies. (Natural or man-made).	purposes (as resting areas for ducks, for example). Depending on the water feeding them and where it flows from, these water bodies can be classified as: > Ponds: freshwater inland water bodies. Fed mainly by their rainwater catchment area. > Lagoons: coastal water bodies separated from the sea by a lido or dune cordon. Lagoon water may be brackish (temporarily or permanently in contact with the sea through a channel) or fresh (completely cut off from the sea and fed by surface run-off, rivers or groundwater). The shallow water promotes the development of both
Pool or other small water body	Small stretches of shallow, stagnant water (maximum 5000 m ² , up to 2 m in depth). The whole water column receives sunlight and plants can take root anywhere on the bottom. (Natural or man-made).	aquatic and amphibious vegetation (the latter being able to survive on dry land). These habitats often support highly productive flora and fauna. This category comprises all small shallow pools of water, including hoof prints, water barrels, drinking troughs and puddles, even if there is no aquatic vegetation due to their ephemeral nature or artificial construction. They are fed by rainwater, surface run-off or seepage. Pools can therefore be affected by climatic variations and dry up in the summer, especially in the Mediterranean zone. Ruts can host certain species but should preferably not be allowed to form as this is not compatible with good soil management
Peat bog	Wetland habitat where the specific ecology has produced peat soil (organic matter that cannot break down well because of the asphyxiating conditions caused by the permanent presence of stagnant or very slow-moving water).	Peat bogs are highly diverse. Some are acid, others alkaline (also referred to as "low marshes"), and characterised by very different plant associations. Acid peat bogs are dominated by <i>Sphagnum</i> (indicators of cold climatic episodes in the past) and carnivorous plants, and alkaline bogs by <i>Carex</i> . Large peat bogs may be associated with streams or rivers and may contain stretches of water. They are also feeding, resting and reproduction zones for animal species seeking an undisturbed environment.
Marshy area	A wetland area where the soil is permanently saturated and often covered with a layer of stagnant water that does not form peat. The water level varies but there is always enough to support hydrophilic vegetation.	Marshes vary widely in size, occupying hollows and gently sloping land, especially in marshland regions. Marshes are often associated with springs, streams and water bodies of various types.
Sea or ocean	Salt water body.	The forest is rarely in direct contact with water, but some forests can be very close to water, especially on steep coastal slopes and rocky shores.

A.4. ROCKY HABITATS

TYPOLOGY OF ROCKY HABITATS										
Types of rocky habitats	How to identify them?	Comments								
Cliff or rock wall higher than that of adult trees	Sub-vertical rock wall several dozen metres in height, always higher than that of adult trees.	Composite habitat due to its large size. Large temperature differences on unshaded sections, very dry conditions due to wind pressure and lack of standing water.								
Rock wall smaller than that of adult trees	Rocky wall or ledge of low height (smaller than that of adult trees).	Composite habitat rich in varied micro-reliefs, characterized by shady and cool conditions due to the presence of trees (at least in the adult stage).								
Rock slab	An extensive sub-horizontal rocky outcrop	The horizontal surface facilitates: > Development of lithosols favourable to plant life; > Formation of small temporary pools of water.								
Lapiaz or large fresh fracture	A carbonate rock surface with regular fissures of varying depths carved out by dissolution of the rock. This type also includes single large, deep fractures across a slab up to several metres in length.	A composite habitat made up of a rock slab or block with fissures where specific climatic and light conditions create a cool, moist and shadowy environment.								
Cave or chasm	Only the opening is visible.	 Very specific microclimatic and light conditions: > Constant temperature and humidity; > Light decreases away from the opening, sometimes to the point of complete darkness. 								
Unstable scree	An unstable accumulation of stones and rocks.	Little or no decomposing organic matter. Instability maintained by: > Shifting, e.g. by a passing large mammal; > Impacts of falling rock (e.g. from an unstable cliff).								

TYPOLOGY OF ROCKY HABITATS										
Types of rocky habitats	How to identify them?	Comments								
Stable rock pile	An accumulation of stabilised stones and rock, either natural (stable scree) or artificial (piled-up stones, stone wall or ruin).	between the rocks and boulders, presence of decomposing organic matter or fine soil in smalle proportions than the rock and sometimes in onl small quantities. In scree, all the communicating micro-cavities make up a particular habitat referred to as a "shallow subterranean habitat", which hosts highly specialized arthropods								
e P.G.										
Pile of boulders	A pile of very large boulders (> 2 m).	Large empty spaces between boulders.								
© NG.		Frequently produce cold humid conditions between boulders.								
Large rocks or rock outcrops other than slab or lapiaz	These are medium-sized rock elements: > Large blocks (from 20 cm to 2 m in height, covering a significant surface area); > Rock outcrop of the underlying rock, that does not form a slab or lapiaz.	Medium rocks, moderately composite, but when many are present in a woodland, they offer habitats in different situations that are particularly attractive to invertebrates and reptiles								
Outcropping pebble bank	An accumulation of pebbles in the floodplain of a river (all stony habitats present in the riverbed itself are included in the aquatic habitat category).	Pebbles that may shift when a river is in flood. The pebbles are often partly covered with vegetation, but only sparsely vegetated deposits are included in this type.								
Deposit of fine sediments, sparsely vegetated	These may be fine sediment deposits: > In the floodplain of a river (the rocky habitats in the stream bed are integrated into the aquatic habitats), deposited during major floods; > In the form of a dune in the littoral zone.	These deposits are gradually vegetated and only sparsely vegetated deposits are included in this type.								
Loose vertical bank or wall of loose material, sparsely vegetated	Unlike the rock walls, these walls are made of loose materials, but of sufficient cohesion to be subvertical. They are found: > On the banks of rivers, > Or on heavily eroded sedimentary materials.	Only deposits that are sparsely vegetated are included in this type. These walls are sufficiently loose to allow digging by birds (Bank swallows, Common kingfishers, etc.), insects, etc.								

Typology of aquatic and rocky habitats from: Emberger C., Larrieu, L., Rotiel S., Gonin, P.: 2023. *Ten key factors for species diversity in forests. Understanding the Index of Biodiversity Potential (IBP).* 2nd edition. Paris: CNPF-IDF, 2023 Photos: L.L.: Laurent Larrieu; N.G.: Nicolas Gouix; P.G.: Pierre Gonin

A.5. COMBINED FIELD SAMPLING TABLE

STAND DATA SHEET Complete a stand data sheet for each stand and a plot data sheet for each plot t maturity indicators and IBP factors taken at the plot scale, aggregate stand-sca	sampled. The stand data sheet is used to reco ale data, and the assessment score. A reconr	rd identification data, samples, stand-sca aissance of the stand is necessary prior	le indicators and factors, supplementary to data collection, including locating po	information, and the calculated tential sites for carrying out the				
survey of the plots. STAND Task/project: task or project for which the sample is taken. Re located. Province: province in which the stand is located. Municipality: municipality: municipality: municipality.	estricted data: Is the data sensitive for publicat cipality where the stand is located. Property: I	ion purposes? Name: stand name. Au ype of property, public or private. Owne	utonomous community: autonomous c er: in the case of public property, indicat	community in which the stand is e the owner. Area: area of the				
Task/project			Rest	ricted data				
Name	Regio	n						
County	Munic	ipality						
Property Dublic Drivate Owner	L		A	rea ha				
SAMPLING Date: sample date. Team: names of the team members p	erforming the sample (at least, the name of the	field team supervisor, or the contractor).						
Date// Team								
HABITAT CORINE/LPEHT: code and name of the most representative according to the Habitats Directive, that is most representative of the stand (see form the canopy of the stand. \Rightarrow \$p pp /ac: code (see list) and/or name of main in each plot weighted by its basal area.	forest habitat of the stand based on the domin list). Biogeographic region: region in which th and companion species respectively. CC: can	ant species (see list). Community inte ne stand is located according to the Habit ppy cover (in %). Ho: dominant height, ii	erest: code and name of the forest habit lats Directive. Main tree species: the tw n metres, calculated as the mean of the	tat of community interest (HCI), wo main native tree species that dominant heights of the species				
Community interest code/name 9								
Biogeographic region: Alpine				Ho				
		Sp ac code/name		HOM				
Species: code and/or name of tree species at any stage of development and po Species code/name:	pulation level, with Ht > 0.5 m, with emphasis o	n those found between plots.						
FOREST LIFE CYCLE PHASES RB Phase: Indicate the forest life cycle phases present in the stand, provided the phases, indicate in the corresponding box whether the space has opened up na	FOREST LIFE CYCLE PHASES IN Phase: Indicate the forest life cycle phases present in the stand, provided the occupied area is at least 200 m ² (8 m radius), with the exception of the regeneration phase, which can be 100 m ² (6 m radius). In the clear and regeneration phases, indicate in the corresponding box whether the space has opened up naturally, due to the fail of trees (dead or otherwise), or due to felling and/or harvesting.							
Gap: Regeneration:	elling		Maturation	Senescence				
Aquatic habitats IBP	Rocky habitats	an ondiv)						
Characteristic natural or artificial aquatic features, permanent or not (see annow). Characteristic nocky landscape habitals (see appendix). Springs or seepage Cliff or rock wall higher than that of adult trees Small stream, unmaintained humid ditch or small channel (width < 1 m)								
FOREST CONTINUITY IBP RB 1956-57 Ortho: proportion of the stand without trees in the 1956-57 orthophotog activity on the land before 1945. Forest continuity: evidence of continuous fo the area. Historical documents: review of documents indicating the age of the	raphs 1945 Ortho: proportion of the stand with rest cover on land observed to be treeless in 19 e forest.	but trees in the 1945 orthophotographs F 45. Reforestation disturbance: eviden	Forest discontinuity: the extent to which ice of soil movement activity for the purp	is there evidence of agricultural oses of reforestation throughout				
1956-57 Ortho, % tree cover: 91-100%		51-75% 26-50	0% 🗌 11-25%	0-10%				
1945 Ortho: 100% treeless	□ partial tree cover		er before 1945					
Forest discontinuity (walls, terracing):	Entire plot └── Localised └── L e edges of former pastures └── I ghing └── Tilling between rows [❑ Not terraced Rocky outcrops that still hav ❑ Stump removal	ve trees 🗌 Other evidence	1				
Historical documents:		•						
ADDITIONAL INFORMATION IBP RB Other species of accompanying flora: selection of key flora that are indicative for the assessment, e.g. if deadwood or tree microhabilats are mostly of the s	of the habitat Habitats of Community Intere ame species, if there are islands of senescence	st: other HCI identified in the stand, wheth , if there are coppices of different comp	her forested or not. Other relevant info osition, etc.) and/or any other stand info	prmation: additional information				

Other accompanying flora species:														
Habitats of Community Interest:														
Othe	Other relevant information:													
Docu	ments:													
INDIC each inc	CATORS AND FACTORS ● For each plot ⇒ Field dat licator, score or variable for the stand. ● Score: score based on the c	ta for the co classification	rrespondin threshold	g indicator RE s in each case	, facto	r 📴 or variable	e (radius and	l area). For the	e stand ⇒	e Figure	: aggregate f	figure for	Star	nd ▼
Dist	• •	Plot ►	1	2	3	4	5	6	7	8	9	10	•Data	Score
Plot	adius m													
Plot														
RB	Prese Species n species					1	1				1	1		
	Basal area m²/ha													
	Volume of trees m³/ha													
	Tree strata n strata													
ORS	Diameter classes n dc			1 1		1	1	1 1			1	1		
CAT	Exceptional trees (DBH ≥ ED) trees/ha													
IN	Abundance of standing deadwood m ³ /ha													
Ϋ́	Abundance of lying deadwood m ³ /ha													
MATUR	Total deadwood abundance m3/ha													
	Proportion of deadwood %													
	Tree microhabitats n types													
	Silvogenetic phases score													
	Forest continuity over time score													
IBP	A: Native tree species n genera													
	B: Vertical vegetation structure n strata													
	C: Standing deadward : "	LDW												
ŝ		MDW												
10		LDW												
FAC	D. Lying deadwood pieces/na	MDW												
AND	E: Very large living trees (VLT												
ST	E. very large living trees (trees/ha)	LT												
	F: Live trees bearing microhabitats trees/ha													
	G: Flower-rich open areas % surface													
۲ ۳	H: Forest continuity over time age of the forest							İ						
NTE CTO	I: Aquatic habitats number of types													
SĂ	J: Rocky habitats number of types													
SAM The san plots of 17.8 m i	PLING AREA AND NUMBER AND SIZE C npled area must represent between 15% and 50% of the stand area and any radius as necessary for the entire sampled area to meet the require n all cases.	OF PLC d be at least ements ©. 1	DTS one hecta Sampling la	ne O. It is hig arger plots is c	hly reco only app	ommended to ca propriate in very	rry out the sa small stands	ampling on circu that are not ste	ular plots. T eeply slopin	his is the g g and/or w	juidance con ith an abund	tained in the lant understo	protocol. Sur ry. The minin	vey as many num radius is
© Sa propo	Minimum sample (ha) < 7	9-24 25-3 4 5 12 15	0 31-36 6 18	37-42 43- 7 8 21 24	48 @ 3 1 4	Number of plo to sample 1 ha radius and	ots needed based on area	Number (n) Radius (m) Area (ha)	1 56.4 3 1.00 0	2 3 9.9 32.6 .50 0.33	4 5 28.2 25 0.25 0.2	6 .2 23.0 2 20 0.17 0	7 8 1.3 19.9 .14 0.13	9 10 18.8 17.8 0.11 0.10

STANDING D Plot radius \Rightarrow From For each species \Rightarrow	EADWOOD 17.8 to 56.4 m: samp Species: code and/or	RB IB	P ding dead tree ies name. Ht	s of any s : height in	pecies, in any metres. DBI	state of d	lecay of Ht ≥ 1 er at breast he	m and DE	8H ≥ 17.5 cm, [,] . If Ht < 1.30 m	whole tree I, it is not r	es and snags. necessary.						
Species	code/name	Htm	DBH cm	Htm	DBH cm	Ht m	DBH cm	Htm	DBH cm	Ht m	DBH cm	Ht m	DBH cm	Ht m	DBH cm	Ht m	DBH cm
																<u> </u>	
LYING DEAD	From 17.8 to 56.4	BP m: samp	le of pieces w	hole or pa	rt thereof of c	leadwood	of any specie	s in any st	ate of decav w	ith Iα≥1	m and diamet	er(Dm)≥	17.5 cm				
From Tree to 90.4 m: sample of pieces, whole of part mereor, or oeadwood of any species, in any state of decay with Lg ≥ 1 m and diameter (Um) ≥ 17.5 cm. For each species ⇒ Species: code and/or species name. Lg: length of the piece, or part thereof, with Dm ≥ 17.5 cm. Dm: diameter of the piece at the midpoint of its length, in cm.																	
Species	code/name	Lgn	n Dm cm	Lg n	n Dm cn	Lg	m Dm c	n Lg	m Dm cn	n Lg	m Dm cr	Lg	m Dm c	m Lg	m Dm cr	n Lg m	Dm cm
																1	-
																+	
TREE OTRA		i		i	1	i	i	_i		_i	1	i		i			
IREE SIRA Plot radius ⇒	15 m: area covere	ed by livin	g trees, visual	ised in ter	ms of four stra	ata of equ	al height up ti	the domi	nant height Th	e emergei	nt stratum (5)	refers to g	enerally isola	ited trees v	whose height is	s greater tha	in that of the
CC ≥ 20%: indicate if	general canopy. the stratum's CC is gro	eater than	20%.														
								1		-	-		_				
		-			-									6			
		1 -11			-4 11-4	11			114 1	- 11		1-4-11	14	11		11-4	n -
CC ≥ 20%:	└ 1 (0 <	≍ Ht ≤ (/4)		2 (¼ < ⊢	lt ≤ ½)		□3(½ < Ht ≤	3⁄4)		4 (¾ <	: Ht ≤ H₀)	□ 5	(H _{em.} >	H₀)
VERTICAL V	EGETATION	STRI	JCTURE	IBP													
Plot radius \Rightarrow CC \ge 20%: indicate if	the stratum's CC is gro	eater than	age of woody 20%.	and herba	ceous vegeta	uon strata	by neight.										
cc > 000/.	Herbaceo	us and	semi-		Very low	woody		Low w	oodv vea	etation		/lid-hei	ght wood	у	🗆 Tall w	oodv ver	aetation
UU ≥ 20%:	woody ve	egetatio	on		(Ht < 1.5)	on m)		(H	t: 1.5-5 m)			Veget (Ht: 5-	ation 15 m		(lt > 15 m)	,
			1				I				1	,					
Plot radius ⇒					From 17.8	to 56.4 m	n: surface area	of open s	paces with per	manent or	r temporary flo	wering veç	getation.				
Openings or c	learings: surf	pe of ope	n space. % fl m² ¥	owering s	% flower	ina sr	ecies =	n spaces c	m ² m ²	wering spo	ecies. Lengti	i: length o	i iorest edge	spaces in r	openinas	or cleari	nas +
Low dense are	eas: surface	m	2 X	% flo	wering s	pecies	=	m ²						low-de	nsity area	s +	
Forest edges	orest edges: length $m \times 2 m$ width = $m^2 \times \frac{9}{10}$ flowering species = m^2																

PLOT Complete a	DATA Sheet	SHEET for each plot s	• sampled. It is us	ed to record fie	eld data samp	oled at plot so	ale using eithe	er Red Bosq	ues (RB)	or IBP standa	ırds. Plot	s should	be located o	on sites co	ontainin	g features th	at denote r	nature fo	rests and/or
are importan area that nee	t for biodiversit eds to be samp	ty (deadwood, led (see tables	large and/or ex in the stand da	ceptional trees, ita sheet). This	, etc.). Plot s size correspo	sizes and dis onds to all ele	tances are me ements, except	asured on tl t for live tree	he horizor es of DC 2	ital plane, not 0 and 25 and	t following the tree s	the slop trata.	e of the land	l. It shou	ld be no	oted that the	size of the	plot dep	ends on the
PLOT	relative numeri	ical identifier fo	or the plot (e.g.,	1/5, 2/5, 3/5…)	. Code: plot	t code, if diffe	rent from the r	number. St	tand: nam	e and/or code	e of the sta	and to wh	ich the plot I	belongs. C	Coordin	ates: geogr	aphic coord	inates (a	lways in
Number			ode:	netres and indic	cating the zor	stand [.]	t from 31. Si te	e quality: so	ee aocum	ent Pique, M.	et al., 201	14. TIPOIC	igies foresta	is arbrade	is. Urg	EST. UPF.	Jencat.		
Coordin	ates ETRS	:89'	Zone	UTM	X m			UTM v	m				Site	quali	tv·⊡		вП	2	
SAMDI	SAMPLE																		
Date: sample minimum sar	▪ ⊑ e date. Team: mpling area rec	: names of the quired. ***	team members	performing the	sample (at le	east, the nam	e of the field to	eam supervi	isor, or the	contractor).	Plot radi	ius: plot :	size, in metro	es, depen	ding on	the number	of plots in f	he stand	and the
Date:	_//		Team:																
Radius	of the plo	t m:	56.4	3	9.9	32.6	2	8.2	25	.2 🗌	23.0		21.3] 19.9] 18.8] 17.8
LIVE T	REES RB	IBP																	
Plot radius	⇒ 1 ci	0 m: count of a m) and DC 25.	all trees, native	or not, in DC 20) (DBH ≥ 17.5	5 From 17.8 species o	to 56.4 m: co f DBH < 17.5 c	ount of all tre cm and Ht ≥	es, native 50 cm. 0	or not, in DC CC of native to	30 to DC	55, and i s (Ht ≥ 5	measuremer 0 cm).	nt of the D	BH of a	II trees with	DBH ≥ 57.9	5 cm. O	ther
For each sp 20→DC 55:	ecies ⇒ Spec number of tree	ies: code and/ es in each diar	or tree species neter class pres	name (see list). ent (take into a	. Ho: domina account the p	ant height, av	verage of the th line with DC).	nree largest DBH ≥ 57	standing t '. 5 : DBH c	rees in DBH (f each tree e	of the plot. xceeding	. ED : ex 57.5 cm i	ceptional dia in the variab	ameter, ED le radius p) (cm) = plot. (= 3 x Ho (m) Other speci	(see equiva es: code a	ilence tal and/or na	ble ⊕). DC ime of other
species pres Hom ≤	ent in DC other	than those de	tailed above.	CC: canopy co	Ver fraction fo	or all native ti	ree species. 22 23 2	4 25	26	27 28	29	30 3	1 32	33 34	3	5 36	37 38	39	≥40
ED cm 2 DC	32.5 30 35	37.5 40	42.5 45	47.5 52 50 5	2.5 57.5 55 60	5 62.5 65	67.5 70	72.5	8	0 82.	5 8 5 9	7.5 90	92.5 95	97.5 100	102	2.5 107 5 11	0	112.5 115	117.5 120
RB ⇒			LT DBH 37.5 -	57.5 cm			Exc	ceptional tr	ees DBH	cm > 3 x Ho r	m — — VLT DE	BH ≥ 57.9	5 cm —						→ →
₿₽⇒	Distan	<u> </u>		VLT D	OBH ≥ 37.5 ci	m / Site quali	ty / Species: C	Crataegus, R	Rhamnus,	Phyllerea Arb	utus, Acei	r, Pyrus, I	Prunus, etc.						→
Snec	Plot ra	aius ⇒ Baius ⇒	m FD cm	DC 20	m DC 25	DC 30	DC 35	DC 40	DC	45 DC	50 [DC 55	06.4 m		П	BH > 57	7.5 cm		
opeo	ico codeman			0020	0020	0000	0000	00 40		+0 00	<u> </u>	0000							
0//																	1. 500		
Other s	becies:												Nat	ive CC	: <	50%∟]≥ 50%	0	
TREE N	/IICROH/	ABITATS	RB IBP	f l'an h		E ann with an	-h -f ih- 45 h			(and table II) If these				1-4 <u>-</u> - 4		4	everte d e	- different
Microhabit	s⇒ tat types:1 W	rees. If a tree	has more than a reeding cavitie	one microhabita	at of the same $(\emptyset > 10 \text{ cr})$	e type it is co m or > 30 cm	unted as one.	Lor open ca	vities) I 3	Insect galle	ries and	hore hol	es (Ø > 2 ci	m) I 4 Co	oncavit	e same tree	cm D>10	cm) I 5	Exposed
sapwood (> 30 cm)	S > 600 cm ² or 7. Crown dead	r split bark > 1 wood Ø > 20	cm, W and heig cm and L > 50	pht > 10 cm) 6 cm or Ø > 3 cm	Exposed h with $> 20\%$	eartwood ar	d sapwood (s d) 8. Burrs a	stem breaka	ige Ø > 20 (Ø > 20 c	cm, broken t m) 9. Twia 1	branch at t tangles (v	trunk leve vitches' b	el (S > 600 c room > 50 c	m ² = A4, (m: offshoo	or Ø > 2 ots > 5)	20 cm); split	of W > 1 cr nial fungal	n, D > 10 fruitina	cm and L
> 5 cm) 1 50 cm) 14	1. Ephemeral . Microsoils (c	fungal fruiting crown, at any h	g bodies (Ø > 5 leight) 15. Sap	cm or N > 10) and resin exu	12. Epiphy idates (lengt	tic or paras h> 20 cm) Ø	itic crypto. An D: diameter; S:	nd phanerog surface; L: I	gams (mo length; W	sses, lichens width; N: nur	or lianas mber: D: d	> 20% of lepth.	the trunk; n	nistletoe >	10 ball	s > 20 cm, fe	erns > 5 lea	ves) 13	. Nests (>
Trees: num	ber of trees with	h the correspo	nding microhab	itat type.	Ja ál	· Infect	li sh	te 1		Jun il			1			1ton		le	eld l
abital e	•	11-	- TU	1.2	17		and the second	h			E		0.00			No.			11SK
typ	1 Cavition		3 Incost	to (m)	19	6 Evro	IF here	1	শ্ব				2	1				1	14
Ē	made by woodpeckers	2. Cav. organic matte	galleries and bore holes	4. Concavities	5. Exposed sapwood	d heartwo	ood 7. Dead	wood 8.Bi wn ca	urrs and ankers	9. Twig tangles	10. Per fun	ennial 1 ⁻ gi	1. Ephemera fungi	al 12. Epip or para	ohytes asites	13. Nests	14. Micr	osoils re	5. Sap and sin exudates
Trees n											-								

A.6. LIST OF CORINE/LPEHT HABITATS

Adapted habitat list from the Spanish Standard List of Terrestrial Habitats. The primarily Mediterranean (MED) formations are indicated (x). These include variants mixed with other species and reforested habitats.

CODE	NAME	MED
41	Deciduous broad-leaf forests	
41.1	Beech (dominated by Fagus sylvatica)	
41.3	Ash (<i>Fraxinus</i> excelsior)	
41.5&1	Acidophilic oak (Quercus petraea)	
41.5&2	Acidophilic oak (Quercus robur)	
41.6	Pyrenean oak (Quercus pyrenaica)	X
41.7&1	Downy oak (Quercus humilis or hybrids)	X
41.7&2	Portuguese oak (Quercus faginea s.l.)	X
41.7&3	Algerian oak (Quercus canariensis)	X
41.83	Maple (<i>Acer</i> spp.)	X
41.84	(Meso-)supramediterranean forests with abundant linden trees (Tilia platyphyllos)	
41.85	European nettle tree (<i>Celtis australis</i>)	
41.86	Non-riparian forests of Fraxinus angustifolia or F. ornus, sometimes with oak or holm oak	X
41.9	Chestnut (forests dominated by Castanea sativa)	X
41.A	European hornbeam (<i>Carpinus betulus</i>)	
41.B	Birch (excluding riparian or marshland)	
41.D	Aspen-dominated forests (Populus tremula)	
41.E	Rowan (Sorbus aucuparia)	
42	Coniferous forests	
42.&1	Silver fir (<i>Abies alba</i>)	
42.19	Spanish fir (<i>Abies pinsapo</i>)	
42.4	Mountain pine (Pinus uncinata)	
42.5	Scots pine (Pinus sylvestris)	
42.6	Austrian pine (<i>Pinus nigra</i> s.l.)	X
42.8&1	Maritime pine (Pinus pinaster)	X
42.83	Stone pine (<i>Pinus pinea</i>), natural or semi-natural groves	X
42.84	Aleppo pine (Pinus halepensis)	Х
42.9	Canary Island pine (Pinus canariensis)	
42.A2	Spanish juniper (<i>Juniperus thurifera</i>)	
42.A6	Tetraclinis (Tetraclinis articulata)	
42.A7	Common yew (Taxus baccata)	
42.A81	Canary Islands juniper (Juniperus cedrus)	

42.A9	Cade juniper (<i>Juniperus oxycedrus</i> s.l.)	
42.AA	Phoenicean juniper (exceptional formations of Juniperus phoenicea)	
44	Woods and other forest formations on riversides or wetlands	
44.1	Alder	X
44.&1	Poplar	X
44.&3	Riverbank willow and bitter willow (<i>Salix</i> ssp.)	X
44.35	Black poplar (Populus nigra), native to northern Iberian Peninsula	
44.62	Mediterranean riverbank field elm (Ulmus minor)	X
44.63	Mediterranean riverbank narrow-leaved ash (Fraxinus angustifolia)	X
44.813	Tamarisk (riverside formations dominated by Tamarix spp.)	Х
45	Sclerophyll and laurophyll	
45 45.11	Sclerophyll and laurophyll Wild olive (<i>Olea europaea subsp. sylvestris</i>)	
45 45.11 45.12	Sclerophyll and laurophyll Wild olive (Olea europaea subsp. sylvestris) Carob (Ceratonia siliqua)	X
45 45.11 45.12 45.2	Sclerophyll and laurophyllWild olive (Olea europaea subsp. sylvestris)Carob (Ceratonia siliqua)Cork oak (Quercus suber)	X X
45 .11 45.12 45.2 45.3	Sclerophyll and laurophyllWild olive (Olea europaea subsp. sylvestris)Carob (Ceratonia siliqua)Cork oak (Quercus suber)Evergreen oak and holm oak (Quercus ilex or Q. rotundifolia)	X X X
45 .11 45.12 45.2 45.3 45.3 45.6	Sclerophyll and laurophyllWild olive (Olea europaea subsp. sylvestris)Carob (Ceratonia siliqua)Cork oak (Quercus suber)Evergreen oak and holm oak (Quercus ilex or Q. rotundifolia)Macaronesian laurel forests	X X X
45 .11 45.12 45.2 45.3 45.6 45.7	Sclerophyll and laurophyllWild olive (Olea europaea subsp. sylvestris)Carob (Ceratonia siliqua)Cork oak (Quercus suber)Evergreen oak and holm oak (Quercus ilex or Q. rotundifolia)Macaronesian laurel forestsPalm groves	X X X

A.7. LIST OF NATIVE TREE SPECIES

Code and name

224 Apollonias barbujana

100	Abies alba	225	Arbutus canariensis	217	Crataegus sp.
105	Abies pinsapo	73	Arbutus unedo	109	Cupressus lusitanica
219	Abies sp.	10	Betula pendula	110	Cupressus sempervirens
1	Acer campestre	11	Betula pubescens	999	Desconocido
2	Acer monspessulanum	212	Betula sp.	227	Dracaena draco
3	Acer opalus	88	Betula tortuosa	79	Erica manipuliflora
4	Acer platanoides	13	Carpinus betulus	20	Fagus sylvatica
5	Acer pseudoplatanus	15	Castanea sativa	228	Ficus carica
215	Acer sp.	226	Celtis australis	21	Fraxinus angustifolia
6	Alnus cordata	75	Ceratonia siliqua	22	Fraxinus excelsior
7	Alnus glutinosa	76	Cercis siliquastrum	23	Fraxinus ornus
216	Alnus sp.	16	Corylus avellana	24	llex aquifolium

91 *llex canariensis*

90 Crataegus monogyna

26	Juglans regia
150	Juniperus cedrus
111	Juniperus communis
112	Juniperus oxycedrus
113	Juniperus phoenicea
114	Juniperus sabina
115	Juniperus thurifera
218	Larix sp.
92	Laurus canariensis
80	Laurus nobilis
27	Malus domestica
93	Myrica faya
232	Myrica rivas-martinezii
233	Ocotea phoetens
28	Olea europaea
199	Otras coníferas
99	Otras planifolias
234	Persea indica
82	Phillvrea latifolia
225	,
235	Phoenix canariensis
235 83	Phoenix canariensis Phyllyrea angustifolia
235 83 237	Phoenix canariensis Phyllyrea angustifolia Picconia excelsa
235 83 237 122	Phoenix canariensis Phyllyrea angustifolia Picconia excelsa Pinus canariensis
235 83 237 122 125	Phoenix canariensis Phyllyrea angustifolia Picconia excelsa Pinus canariensis Pinus halepensis
 235 83 237 122 125 128 	Phoenix canariensis Phyllyrea angustifolia Picconia excelsa Pinus canariensis Pinus halepensis Pinus mugo (P. montana)
 235 83 237 122 125 128 129 	Phoenix canariensis Phyllyrea angustifolia Picconia excelsa Pinus canariensis Pinus halepensis Pinus mugo (P. montana) Pinus nigra
 235 83 237 122 125 128 129 130 	Phoenix canariensis Phyllyrea angustifolia Picconia excelsa Pinus canariensis Pinus halepensis Pinus mugo (P. montana) Pinus nigra Pinus pinaster

134	Pinus sylvestris
135	Pinus uncinata
85	Pistacia terebinthus
239	Pleiomeris canariensis
31	Populus alba
34	Populus nigra
211	Populus sp.
35	Populus tremula
36	Prunus avium
37	Prunus dulcis
38	Prunus padus
40	Pyrus communis
240	Quercus canariensis
42	Quercus coccifera
43	Quercus faginea
45	Quercus fruticosa (Q. lusitanica)
49	Quercus humilis
46	Quercus ilex
47	Quercus macrolepis
48	Quercus petraea
50	Quercus pyrenaica
51	Quercus robur
52	Quercus rotundifolia
54	Quercus suber
87	Rhamnus alaternus
57	Salix alba
24	Salix atrocinerea

- 58 Salix caprea
- 59 Salix cinerea

- 60 Salix eleagnos
- 61 Salix fragilis
- 62 Salix sp.
- 242 Sambucus nigra
- 243 Sideroxylon mirmulano
- 63 Sorbus aria
- 64 Sorbus aucuparia
- 65 Sorbus domestica
- 66 Sorbus torminalis
- 67 Tamarix africana
- 670 Tamarix sp.
- 137 Taxus baccata
- 245 Tetraclinis articulata
- 68 Tilia cordata
- 69 Tilia platyphyllos
- 210 Tilia sp.
- 70 Ulmus glabra
- 72 Ulmus minor
- 213 Ulmus sp.
- 247 Visnea mocanera

A.8. CODE LIST OF HABITATS OF COMMUNITY INTEREST (HCI) AND PRIORITY HABITATS OF COMMUNITY INTEREST (PHCI)

The Habitats Directive defines natural Habitats of Community Interest as natural or semi-natural terrestrial or aquatic areas that, within the territory of the Member States of the EU: a) are in danger of disappearance in their natural range; b) have a small natural range following their regression or by reason of their intrinsically restricted area; c) present outstanding examples of typical characteristics of one or more of the European Union's biogeographical regions. Among them, priority natural habitat types are those that are in danger of disappearance within the territory of the European Union and for the conservation of which the Community has particular responsibility. In the table, these are indicated with an * after the habitat code.

CODE	NAME
9120	Atlantic acidophilous beech forests with llex and sometimes also Taxus in the shrublayer (<i>Quercion robori-petraeae</i> or <i>Ilici-Fagenion</i>)
9130	Asperulo-Fagetum beech forests
9140	Medio-European subalpine beech woods with Acer and Rumex arifolius
9150	Medio-European limestone beech forests of the Cephalanthero-Fagion
9160	Sub-Atlantic and medio-European oak or oak-hornbeam forests of the Carpinion betuli
9180*	Tilio-Acerion forests of slopes, screes and ravines
91B0	Thermophilous Fraxinus angustifolia woods
91D0*	Bog woodland
91E0*	Alluvial forests with Alnus glutinosa and Fraxinus excelsior (Alno-Padion, Alnion incanae, Salicion albae)
9230	Galicio-Portuguese oak woods with Quercus robur and Quercus pyrenaica
9240	Quercus faginea and Quercus canariensis Iberian woods
9260	Castanea sativa woods
92A0	Salix alba and Populus alba galleries
92B0	Riparian formations on intermittent Mediterranean water courses with <i>Rhododendron ponticum</i> , <i>Salix</i> and others
92D0	Southern riparian galleries and thickets (Nerio-Tamaricetea and Securinegion tinctoriae)

Olea and Ceratonia forests
Quercus suber forests
Quercus ilex and Quercus rotundifolia forests
Macaronesian laurel forests (Laurus, Ocotea)
Palm groves of <i>Phoenix</i>
Forests of <i>llex aquifolium</i>
Subalpine and montane Pinus uncinata forests (* if on gypsum or limestone)
Subalpine and montane Pinus uncinata forests (* if on gypsum or limestone)
Abies pinsapo forests
(Sub-) Mediterranean pine forests with endemic black pines
Mediterranean pine forests with endemic Mesogean pines
Canary Island endemic pine forests
Endemic forests with Juniperus spp
Tetraclinis articulata forests
Mediterranean Taxus baccata woods

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